



UNIVERSITI PUTRA MALAYSIA

***INFLUENCE OF POLYCYCLIC AROMATIC HYDROCARBON AND
GENETIC POLYMORPHISMS (CYP1A1, GSTM1, GSTT1) ON PYRENE
METABOLITE AND DNA DAMAGE AMONG CHILDREN IN THE KLANG
VALLEY, MALAYSIA***

NUR HAZIRAH BINTI HISAMUDDIN

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**Thesis Submitted to the School of Graduate Studies, Universiti Putra Malaysia in
Fulfilment of the Requirements for the Degree of Doctor of Philosophy**

June 2022

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Abstract of thesis presented to the Senate of Universiti Putra Malaysia in fulfilment of the requirement for the degree of Doctor of Philosophy

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June 2022

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Polycyclic Aromatic Hydrocarbon (PAHs) are hazardous particulate matter bound-organic compound found in urban atmospheric environment. Exposure to PAHs is of great concern to public health especially for children, due to their ubiquitous presence in the environment and well-recognised carcinogenicity, teratogenicity and genotoxicity to human health. This study aimed to assess the association of particulate PAHs exposure on biomarker of PAHs exposure, urinary 1-hydroxypyrene (1-OHP) and DNA damage among children in the vicinity of heavy traffic areas in the Klang Valley. In addition, this study investigated the effects of genetic polymorphisms of PAHs metabolising enzymes on the concentrations of pyrene metabolite, the urinary 1-OHP level, and on DNA damage in children. This cross-sectional comparative study was conducted at eight primary schools in high traffic (HT) and low traffic (LT) areas in Klang Valley, Malaysia. A validated questionnaire on socio-demographic, health status, personal exposure and dietary intake was distributed to randomly selected children. The indoor and outdoor samples of particulate matter PM_{2.5}-bound PAHs were collected using a MiniVol portable air sampler for 24 h. PAHs concentration was quantified using Gas Chromatography-Mass Spectrometry (GC-MS). Source diagnostic ratio was applied to determine the source of PAHs emission. DNA damage in buccal cells were evaluated using Comet Assay, while urinary 1-OHP was assessed using High-Performance Liquid Chromatography (HPLC). Genes encoding three PAHs-metabolising enzymes *CYP1A1*, *GSTT1* and *GSTM1* were studied for polymorphisms using PCR-RFLP and multiplex PCR. The results showed that the total outdoor PAHs was higher (range between 1.36-5.76 ng/m³) than indoor environments (range between 1.25-5.58 ng/m³). The concentration of total PAHs was higher at HT compared to LT area. Based on the source diagnostic ratio, PAHs in the HT area is pyrogenic, mainly from high combustion activity (vehicular emission). Meanwhile, the urinary 1-OHP concentration was significantly higher among children in HT as compared to LT (0.07 vs 0.04 µmol/mol-creatinine). The DNA damage was obviously higher among children in HT as compared to LT. A

significant association were found between environmental PAHs exposure with urinary 1-OHP and DNA damage. It was observed that there was no significant effect of *CYP1A1*, *GSTM1* and *GSTT1* genotypes on urinary 1-OHP and DNA damage levels. The 95th percentile of the incremental lifetime cancer risk (ILCR) using Monte Carlo simulation showed that the inhalation risk for the children in HT and LT area were 2.80×10^{-7} and 1.43×10^{-7} , respectively. After controlling the possible confounders, the result from multiple linear regression showed that PAHs exposure was the most significant factor associated with urinary 1-OHP and DNA damage among children. This study provided evidence that children living near heavy traffic area are more likely to be exposed to environmental PAHs and have a higher risk of DNA damage than children living in low traffic areas. Findings from this study will be able to help in minimising the potential health effect related to traffic related air pollution in urban area by designing a mitigation plan for school.

Keywords: PAHs, children, urinary 1-OHP, DNA damage, genetic polymorphisms

Abstrak tesis yang dikemukakan kepada Senat Universiti Putra Malaysia sebagai memenuhi keperluan untuk ijazah Doktor Falsafah

**PENGARUH HIDROKARBON AROMATIK POLISIKLIK DAN
POLIMORFISME GENETIK (*CYP1A1*, *GSTM1*, *GSTT1*) PADA METABOLIT
PIRENA DAN KEROSAKAN DNA DALAM KALANGAN KANAK-KANAK DI
LEMBAH KLANG, MALAYSIA**

Oleh

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Hidrokarbon Aromatik Polisiklik (HAP) ialah sebatian organik terikat zarah berbahaya yang terdapat dalam persekitaran atmosfera bandar. Pendedahan kepada HAP amat membimbangkan kesihatan awam terutamanya kepada kanak-kanak, kerana kehadirannya di alam sekitar dan sifat karsinogenik, teratogenik dan genotoksik kepada kesihatan manusia. Kajian ini bertujuan untuk menilai hubungan pendedahan HAP partikulat terhadap penanda biologi HAP, urin 1-hidroksipirena (1-OHP) dan kerosakan DNA dalam kalangan kanak-kanak di sekitar kawasan trafik tinggi di Lembah Klang. Sebagai tambahan, kajian ini menilai kesan polimorfisme genetik enzim metabolisme HAP terhadap kepekatan metabolit pirena iaitu urin 1-OHP dan kerosakan DNA pada kanak-kanak. Kajian keratan rentas perbandingan ini dilakukan di lapan buah sekolah rendah di kawasan trafik tinggi (HT) dan kawasan trafik rendah (LT) di Lembah Klang. Borang soal selidik yang telah divalidasi berkaitan sosiodemografi, status kesihatan, pendedahan individu dan pemakanan telah diedarkan kepada kanak-kanak yang dipilih secara rawak. Sampel partikulat PM_{2.5} (dalaman dan luaran) dikumpulkan menggunakan pensampel udara mudah alih MiniVol selama 24 jam. Kepekatan HAP diukur menggunakan Kromatografi Gas Spektroskopi Jisim. Nisbah sumber diagnostik digunakan untuk menentukan punca emisi HAP. Kerosakan DNA pada sel pipi dinilai menggunakan analisis komet, sementara urin 1-OHP dinilai menggunakan Kromatografi Cecair Prestasi Tinggi. Gen enzim metabolisme HAP, *CYP1A1*, *GSTT1* dan *GSTM1* dikaji untuk polimorfisme menggunakan PCR-RFLP dan multipleks PCR. Hasil kajian menunjukkan bahawa kepekatan HAP luaran lebih tinggi (julat antara 1.36-5.76 ng/m³) berbanding kepekatan dalaman (julat antara 1.25-5.58 ng/m³). Kepekatan keseluruhan HAP lebih tinggi di kawasan HT berbanding di kawasan LT. Berdasarkan nisbah sumber diagnostik, HAP di kawasan HT adalah pirogenik, berpunca daripada aktiviti pembakaran tinggi (pelepasan kenderaan). Sementara itu, kepekatan urin 1-OHP adalah lebih tinggi dalam kalangan kanak-kanak di HT berbanding LT (0.07 vs 0.04 µmol/mol-kreatinin). Kerosakan DNA jelas lebih tinggi dalam kalangan kanak-kanak di HT

berbanding LT. Terdapat hubungan yang signifikan antara pendedahan HAP persekitaran dengan kerosakan DNA dan urin 1-OHP. Tiada kesan yang signifikan dari gen *CYP1A1*, *GSTT1* dan *GSTM1* terhadap urin 1-OHP dan kerosakan DNA. Berdasarkan model Risiko Kanser Seumur Hidup (ILCR) menggunakan simulasi Monte Carlo, risiko kanser bagi kanak-kanak di kawasan HT dan LT adalah 2.80×10^{-7} dan 1.43×10^{-7} . Hasil dari Regresi Linier Berganda menunjukkan bahawa pendedahan kepada HAP adalah faktor yang paling signifikan yang berkait dengan urin 1-OHP dan kerosakan DNA, setelah mengawal faktor-faktor pembauran yang berkemungkinan. Kajian ini menunjukkan bahawa kanak-kanak yang tinggal berhampiran trafik tinggi lebih cenderung terdedah kepada HAP dan mempunyai risiko kerosakan DNA yang lebih tinggi berbanding kanak-kanak yang tinggal di kawasan trafik rendah. Penemuan dari kajian ini akan dapat membantu meminimumkan potensi kesan kesihatan dari pencemaran udara berkaitan trafik di kawasan bandar dengan merancang pelan mitigasi strategik untuk sekolah.

Kata kunci: PAHs, kanak-kanak, kerosakan DNA, urin 1-hidroksipirena, polimorfisme genetik

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TABLE OF CONTENTS

	Page
ABSTRACT	i
ABSTRAK	iii
ACKNOWLEDGEMENTS	v
APPROVAL	vi
DECLARATION	viii
LIST OF TABLES	xiv
LIST OF FIGURES	xvii
LIST OF APPENDICES	xix
LIST OF ABBREVIATIONS	xx
 CHAPTER	
 1 INTRODUCTION	 1
1.1 Study Background	1
1.2 Problem Statement	2
1.3 Study Justification	5
1.4 Conceptual Framework	6
1.5 Research Objectives	8
1.5.1 General Objective	8
1.5.2 Specific Objectives	8
1.6 Research Hypothesis	9
1.7 Definition of Variables	9
1.7.1 Conceptual Definition	9
1.7.2 Operational Definition	12
 2 LITERATURE REVIEW	 13
2.1 Particulate Matter	13
2.2 Polycyclic Aromatic Hydrocarbon (PAHs)	14
2.2.1 High and Low Molecular Weight PAHs	16
2.2.2 Source Identification of PAHs	17
2.3 Health Risk Assessment	19
2.4 Children as Susceptible Population	19
2.5 Biomarkers of Effect	20
2.5.1 DNA Damage	20
2.5.2 Comet Assay	21
2.6 Biomarker of Exposure	24
2.6.1 Urinary 1-hydroxypyrene	24
2.7 Metabolic Activation of PAHs	28
2.8 Biomarker of Susceptibility	30
2.8.1 <i>CYP1A1</i>	30
2.8.2 <i>GSTM1</i> and <i>GSTT1</i>	31

3	METHODOLOGY	33
3.1	Study Design	33
3.2	Study Location	34
3.3	Study Sampling	39
3.3.1	Sampling Period	39
3.3.2	Study Population	39
3.3.3	Sampling Frame	39
3.3.4	Sampling unit	39
3.3.5	Sample Size	40
3.3.6	Sampling Method and Subject Recruitment	42
3.4	Questionnaire	43
3.5	Diary of Daily Activities	44
3.6	Environmental Sampling	44
3.6.1	Measurement of Particulate Matter in Schools	45
3.6.2	Measurement of Meteorological Parameters	46
3.6.3	Traffic Count	46
3.6.4	Analysis of PAHs in PM _{2.5} and GC-MS analysis	46
3.6.5	Measurement of PM _{2.5} in Residences	47
3.7	Anthropometry Measurements	48
3.8	Biological Sampling	48
3.8.1	Buccal Cell Collection	48
3.8.2	Urine Collection	49
3.9	Comet Assay	49
3.10	Urinary 1-OHP	50
3.10.1	Sample Preparation and Clean-up	50
3.10.2	High-Performance Liquid Chromatography (HPLC)	51
3.10.3	Creatinine Analysis	51
3.11	Genetic Polymorphism	52
3.11.1	DNA Extraction	52
3.11.2	Genotyping Assay for <i>GSTM1</i> and <i>GSTT1</i>	53
3.11.3	Genotyping Assay for <i>CYP1A1</i>	54
3.12	Health Risk Assessment	56
3.13	Statistical Analyses	57
3.14	Study Ethics	58
3.15	Quality Control	58
3.15.1	Questionnaire	58
3.15.2	Air Sampling	58
3.15.3	Biological Sampling	59
3.15.4	PAHs Extraction and GC-MS Analysis	59
3.15.5	Urine Extraction and HPLC Analysis	59
3.15.6	Comet Assay	60
3.15.7	PCR	60
4	RESULTS AND DISCUSSION	61
4.1	Characteristics of Respondents	61
4.1.1	Response Rate	61
4.1.2	Sociodemographic and Socioeconomic Background of Respondents	62
4.1.3	Body Mass Index (BMI)	64
4.1.4	Background of Residences	65
4.1.5	Home Environment	66

4.1.6	Dietary Intake	67
4.1.7	Respondents' Daily Activities	68
4.2	Respondents' Traffic Exposure	70
4.3	Concentration of PM and PAHs at Schools	74
4.3.1	Indoor and Outdoor PM Concentrations in the Schools (6 h)	74
4.3.2	Indoor and Outdoor PM _{2.5} (24 h)	81
4.3.3	Distribution of PAHs Species	82
4.3.4	Indoor and Outdoor PAHs	88
4.3.5	Distribution of PAHs Based on Number of Rings	89
4.3.6	Source Diagnostic Ratio	90
4.4	Health Risk Assessment	92
4.5	Concentrations of PM _{2.5} at Residences	98
4.6	Urinary 1-OHP as Biomarker of Exposure	101
4.6.1	Concentration of Urinary 1-OHP	101
4.6.2	Association between PAHs exposure and 1-OHP	109
4.7	DNA Damage as Biomarker of Effect	110
4.7.1	Level of DNA damage	110
4.7.2	Association between PAHs exposure and DNA Damage	117
4.8	Genetic Polymorphism as Biomarker of Susceptibility	118
4.8.1	Distribution of Genotypes	118
4.8.2	Influence of Genetic Polymorphisms of <i>CYP1A1</i> , <i>GSTT1</i> and <i>GSTM1</i> on Urinary 1-OHP	120
4.8.3	Influence of Genetic polymorphisms of <i>CYP1A1</i> , <i>GSTT1</i> and <i>GSTM1</i> on DNA Damage	125
4.9	Correlation between PAHs Exposure and Biomarker of Exposure and Effect	128
4.10	Factors that Influence Excretion of Urinary 1-OHP among Children	130
4.11	Factors that Influence DNA Damage among Children	132
5	SUMMARY, IMPLICATIONS, RECOMMENDATIONS AND CONCLUSION	135
5.1	Summary	135
5.1.1	Measurement of Indoor Particulate Matter (PM ₁ , PM _{2.5} , PM ₁₀) and Physical Parameter	135
5.1.2	Measurement of Outdoor Particulate Matter (PM ₁ , PM _{2.5} , PM ₁₀) and Meteorological Factors	136
5.1.3	24 h Measurement of Indoor and Outdoor PM _{2.5}	136
5.1.4	Distributions of PAHs	136
5.1.5	Source Diagnostic Ratio	137
5.1.6	Exposure Assessment	137
5.1.7	Health Risk Assessment	137
5.1.8	Concentrations of PM _{2.5} at Residences	137
5.1.9	Urinary 1-OHP	137
5.1.10	DNA Damage	138
5.1.11	Genetic Polymorphism	138
5.1.12	Correlation between PAHs and Biomarkers	139
5.1.13	Factor that Influences Urinary 1-OHP	139
5.1.14	Factor that Influences DNA Damage	139

5.2	Conclusion	140
5.3	Strength	140
5.4	Limitations	141
5.5	Recommendations	141
5.5.1	School Management	142
5.5.2	Parents or Guardian	143
5.5.3	Authorities	143
5.5.4	Future Research	144
REFERENCES		145
APPENDICES		177
BIODATA OF STUDENT		197
LIST OF PUBLICATIONS		198



LIST OF TABLES

Table	Page
2.1 Summary of Studies on Ambient PAHs in Klang Valley	16
2.2 Abbreviation of 16 USEPA Priority PAHs, Number of Rings and Phase Distribution	17
2.3 Typical reported values of PAH diagnostic ratio for a source category	18
2.4 Summary of Epidemiological Studies Regarding to PAHs Exposure on DNA Damage among Children Population	23
2.5 Summary of Studies Reporting Urinary 1-OHP Concentrations (μmol/mol of Creatinine) in Children Around the World	26
2.6 Studies Involving the Related Genotypes Influencing Urinary 1-OHP Level in Children	32
3.1 Descriptive Information of the Classroom of each School	38
3.2 The Estimation of Sample Size for Each Studied Variable	40
3.3 The Sample Size of Each Stratum in the HT Group	43
3.4 The Sample Size of Each Stratum in the LT Group	43
3.5 Forward and Reverse Primers Sequence for <i>GSTT1</i> and <i>GSTM1</i> Genes	53
3.6 The Components of Multiplex PCR Mixtures for <i>GSTT1</i> and <i>GSTM1</i>	53
3.7 Multiplex PCR Conditions for <i>GSTT1</i> and <i>GSTM1</i> Genes	54
3.8 Forward and Reverse Primers Sequence for <i>CYP1A1</i> Genes	54
3.9 Components of the PCR Mixtures for the <i>CYP1A1</i> Genes	55
3.10 PCR Conditions for <i>CYP1A1</i> Genes	55
3.11 The Distribution Type of Variables Applied in the Monte Carlo Simulation for schools in HT and LT	57
4.1 Sociodemographic Characteristic of Respondents	63

4.2	Socioeconomic Characteristic of Respondents	64
4.3	Household Income Categories of Respondents	64
4.4	Body Mass Index (BMI) of Respondents	65
4.5	Background of Residences between Two Groups	66
4.6	Sources of Indoor Air Pollutants at Residences	67
4.7	Dietary Intake between Two Groups	68
4.8	Indoor PM Concentrations and Physical Parameters (Temperature, Relative Humidity and Air Velocity) in the Schools (6 h)	76
4.9	Outdoor PM Concentrations and Meteorological Parameters in the Schools (6 h)	77
4.10	Comparison of Indoor and Outdoor PM _{2.5} Concentrations Measured for 24 Hours in HT and LT Schools	81
4.11	Distributions of PAHs species in indoor PM _{2.5} samples (N= 32)	84
4.12	Distributions of PAHs Species in Outdoor PM _{2.5} Samples (N= 32)	85
4.13	Concentrations of PAHs in Indoor and Outdoor Air From Schools Around the World	87
4.14	TEQ of PAHs Compounds in Outdoor Air of Studied Primary Schools	94
4.15	TEQ of PAHs Compounds in Indoor Air of Studied Primary Schools	95
4.16	Indoor PM _{2.5} Concentrations in Residences	99
4.17	Associations between PM _{2.5} Inside Residences and Air Pollutant Sources at Residences	100
4.18	Concentration of Urinary 1-OHP in Children of all Studied Schools (μmol/mol Creatinine)	101
4.19	Comparison on Urinary 1-OHP between Schools in HT and LT Group	101
4.20	Comparison of Urinary 1-OHP Concentration in Previously Published Studies for Children Populations (μmol/mol-Creatinine)	103

4.21	Comparison of Urinary 1-OHP Concentrations ($\mu\text{mol/mol}$ Creatinine) According to Demographic, Indoor and Outdoor Environmental Exposure and Dietary Variables	106
4.22	Association between PAHs Exposure and Urinary 1-OHP	109
4.23	The mean Comparison Tail Moment in Children of all Studied Schools	111
4.24	Comparison of Tail Moment between Schools in HT and LT Group	111
4.25	Comparison of Comet Assay Findings in Previously Published Studies for Children Populations	112
4.26	Comparison of Tail Moment According to Demographic, Indoor and Outdoor Environmental Exposure and Dietary Variables	114
4.27	Association Between PAHs Exposure and DNA Damage	117
4.28	Genotype Distribution of the Studied Population	120
4.29	Excretion of urinary 1-OHP ($\mu\text{mol/mol}$ creatinine) According to <i>CYP1A1</i> , <i>GSTT1</i> and <i>GSTM1</i> Genotypes among Study Group	124
4.30	Tail Moment According to <i>CYP1A1</i> , <i>GSTT1</i> and <i>GSTM1</i> Genotypes among Study Group	127
4.31	Correlation between PAHs Exposure, Urinary 1-OHP and DNA Damage	129
4.32	Relationship between Urinary 1-OHP with PAHs Exposure, Polymorphisms of <i>CYP1A1</i> , <i>GSTT1</i> , <i>GSTM1</i> and Other Risk Factors	131
4.33	Factor that Influence Excretion of Urinary 1-OHP among Children after Controlling all Confounders	131
4.34	Relationship between DNA Damage with PAHs Exposure, Polymorphisms of <i>CYP1A1</i> , <i>GSTT1</i> , <i>GSTM1</i> and other Risk Factors	133
4.35	Factor that Influence DNA Damage among Children after Controlling all Confounders	134

LIST OF FIGURES

Figure	Page
1.1 Exposure-disease Continuum Involving Biomarkers of Susceptibility, Exposure and Effects	2
1.2 Total Motor Vehicles by Type in Malaysia as recorded until 31 Disember 2019	3
1.3 Conceptual Framework	7
2.1 Mechanism of PM-induced Health Effects Associated with Oxidative Stress, Inflammation and Cancer Risk	14
2.2 Structure of 16 USEPA Priority Pollutant PAHs	15
2.3 DNA Adduct Formation in Three Pathways of Metabolic Activation of PAHs (Benzo[a]Pyrene is Used as the Representative of PAHs)	21
2.4 Tail Moment	22
2.5 Chemical Structure of 1-hydroxypyrene	24
2.6 Metabolic activation of PAHs	29
3.1 General Flowchart of Research	33
3.2 Locations of the Selected Primary Schools	35
3.3 Coordinates and Location of Schools in HT Area	36
3.4 Coordinates and Location of Schools in LT Area	37
3.5 a) Outdoor and b) Indoor Air Sampling using MiniVol Air Sampler	46
3.6 Buccal Cell Collection using Swab	49
3.7 Elution of 1-OHP	51
3.8 Calibration Curve for Urinary 1-OHP	60
4.1 Flowchart of Respondents' Selection	62
4.2 Daily Activities of Respondents during a) Weekdays and b) Weekends	69

4.3	Mode of Transportation to School	70
4.4	Number of Respondents Who Reported <100 and ≥ 100 Vehicles Around their Residences on Weekdays and Weekends in (a) HT group (b) LT Group	71
4.5	Perception of Traffic Exposure in Respondents' Residences among (a) HT Group and (b) LT Group	72
4.6	Average Count of Vehicles Passed by the Schools in the a) Morning and b) Afternoon	73
4.7	Average Count of Idling Vehicles for 30 Minutes in the a) Morning and b) Afternoon	74
4.8	Hourly Variations of (a) Indoor and (b) Outdoor PM Concentrations in HT Schools	79
4.9	Hourly Variations of (a) Indoor and (b) Outdoor PM Concentrations in LT Schools	80
4.10	Average Concentration of Indoor and Outdoor PM _{2.5} (24 h)	81
4.11	Indoor to Outdoor (I/O) Ratio of PM _{2.5} Concentration	82
4.12	Indoor to Outdoor (I/O) Ratio of PAHs Concentration	89
4.13	Percentage Distribution of Outdoor and Indoor PAHs based on Number of Rings	90
4.14	The Bivariate Plots of Selected PAHs Based on the Indicator Source of Diagnostic Ratio	92
4.15	Probabilistic Distribution of ILCR for Children in HT and LT	97
4.16	Sensitivity Analyses of Carcinogenic Risk for PAHs Exposure among Children in a) HT and b) LT Groups	98
4.17	Comet Cells Observed under the Fluorescence Microscope (20x Magnification 1) Comet tail 2) Undamaged DNA	110
4.18	Multiplex PCR Amplification of <i>GSTT1</i> and <i>GSTM1</i> Resolved in 3% Agarose gel Electrophoresis in Representative Samples (1-10)	118
4.19	Restriction Fragments for <i>CYP1A1</i> Gene Resolved in 2.5% Agarose Gel Electrophoresis in Representative Samples (1-10)	119
4.20	Illustration of Biotransformation and Elimination of PAHs	123
5.1	Strategies for Reducing Exposure to Traffic Pollution in School	141

LIST OF APPENDICES

Appendix	Page
A Questionnaire	177
B Diary of Daily Activities	184
C Comet Assay	185
D Urinary 1-OHP	186
E PCR	188
F MOE Approval Letter	189
G UPM Ethics Committee Approval Letter	190
H Consent Form	192

LIST OF ABBREVIATIONS

<	Less than
>	More than
≥	At least
%	Per cent
°C	Degree celsius
Σ	Summation
ΣPAHs	Total PAHs
χ^2	Chi square
x g	Times gravity
μL	Microlitre
μm	Micrometre
μM	Micromolar
μg/m ³	Microgram per metre cubic
μmol/mol	Micromol per mol
1-OHP	1-hydroxypyrene
ACP	Acenaphthene
ACY	Acenaphthelyne
ADT	Average daily traffic
ANT	Anthracene
ANOVA	Analysis of variance
AT	Averaging time
ATS	American Thoracic Society

ATSDR	Agency for Toxic Substances and Disease Registry
Avg	Average
BaA	Benzo[a]anthracene
BaP	Benzo[a]pyrene
BeP	Benzo[e]pyrene
BaPeq	BaP-equivalent concentration
BbF	Benzo[b]fluoranthene
BkF	Benzo[k]fluoranthene
BgP	Benzo[g,h,i]perylene
BMI	Body mass index
BW	Body weight
cf	Conversion factor
CI	Confidence interval
cPAHs	Carcinogenic PAHs
CSF	Cancer slope factor
CYP	Cytochrome P450
CYR	Chrysene
<i>CYP1A1</i>	<i>Cytochrome P450 1A1</i>
DhA	Dibenzo(a,h)anthracene
DNA	Deoxyribonucleic acid
DOE	Department of Environment
ED	Exposure duration
EDTA	Ethylenediaminetetraacetic acid
EF	Exposure frequency

ETS	Environmental tobacco smoke
FLR	Fluorene
FLT	Fluoranthene
g/L	Gram per litre
GC-MS	Gas Chromatography-mass spectrometry
GM	Geometric mean
<i>GST</i>	<i>Glutathione S-transferase</i>
<i>GSTM1</i>	<i>Glutathione S-transferase Mu 1</i>
<i>GSTT1</i>	<i>Glutathione S-transferase Theta 1</i>
h	Hours
HMW	High molecular weight
HPLC	High-performance liquid chromatography
HT	High traffic
IARC	International Agency for Research on Cancer
IcP	Indeno(1,2,3-cd)pyrene
ILCR	Incremental lifetime cancer risk
IQR	Interquartile range
IR	Inhalation rate
L/min	Litre per minute
LMW	Low molecular weight
LOD	Limit of detection
LT	Low traffic
M	Molar
m/s	Metre per second

Max	Maximum
Min	Minimum
MLR	Multiple linear regression
MOE	Ministry of Education
N	Total population
n	Number of samples
NAP	Naphthalene
ncPAHs	Non-carcinogenic PAHs
ng/g	Nanogram per gram
ng/m ³	Nanogram per metre cubic
PAHs	Polycyclic aromatic hydrocarbons
PBS	Phosphate buffer solution
PCR	Polymerase chain reaction
PHE	Phenanthrene
ppb	Part per billion
ppm	Part per million
PM	Particulate matter
PM ₁	Particulate matter below 1 micrometre aerodynamic diameter
PM _{2.5}	Particulate matter below 2.5 micrometres aerodynamic diameter
PM ₁₀	Particulate matter below 10 micrometres aerodynamic diameter
PYR	Pyrene
RH	Relative humidity
ROS	Reactive oxygen species

rpm	Rotation per minute
SE	Standard error
SD	Standard deviation
SLR	Simple linear regression
TEQ	Toxicity equivalent concentration
TEF	Toxic equivalent factor
tPAHs	Total PAHs (Carcinogenic and Non-carcinogenic)
USEPA	United States Environmental Protection Agency
V	Volt
WHO	World Health Organization
WS	Wind speed

CHAPTER 1

INTRODUCTION

1.1 Study Background

Urbanisation and economic growth are closely related to increased transportation demand and number of road vehicles in cities (Nagpure et al. 2016, Azhari, Latif, and Mohamed 2018). Pollutants from traffic emissions, also known as traffic-related air pollution (TRAP), are the main contributor to air pollution in Malaysia, especially in urban areas (Awang et al., 2020; Mohd Shafie & Mahmud, 2020). Klang Valley, the most densely populated area in Malaysia, accounts for a large fraction of the anthropogenic sources that have contributed to the worsening of air quality in recent years (Latif et al., 2021). Air quality in Klang Valley frequently has been studied for many years as it is recognised as the central pollutant emission source area in Malaysia, especially particulate matter (PM) pollution (Amil et al., 2016; Juneng et al., 2011; Khan et al., 2015; Latif et al., 2014).

Polycyclic aromatic hydrocarbons (PAHs) are one of the most significant organic groups bound to PM in terms of health risk (Jakovljević et al., 2020). PAHs are ubiquitous, semi-volatile and persistent organic pollutants generated mainly during the incomplete combustion of organic materials (Sulong et al., 2019; Suradi et al., 2021; Urbancova et al., 2020). Due to their resistance to degradation processes, especially when bound to particles, they are subjected to the long-range atmospheric transport and could be detected even in remote areas (Liu et al., 2018). Previous studies have shown the contribution of vehicles emission on PAHs level in urban areas (Bahry et al., 2009; Omar et al., 2006; Omar et al., 2002). The impact of PAHs on the environment and public health are of great concern due to their mutagenic and carcinogenic properties (IARC, 2010, 2021). PAHs congeners are ranked 9th out of 275 hazardous chemicals on the Agency for Toxic Substances and Disease Registry's (ATSDR) priority list, indicating a severe health risk to humans and the environment (ATSDR, 2020). Exposure to PAHs have been associated with various toxicological effects, including immunotoxicity, reproductive and developmental toxicity and carcinogenic and genotoxic effects (Huang et al., 2019). In particular, human exposure to PAHs has been linked to many diseases including cancer, diabetes, cardiovascular disease and hypertension (Fernández et al., 2021).

The application of biomarkers to investigate the effects of PAH exposure helped researchers understand the contribution of environmental exposure to cancer risk, particularly in areas with moderate or low air pollution (Sopian et al., 2021; Suhaimi & Jalaludin, 2015). Biomarkers are divided into three main categories; biomarkers of exposure, effect, and susceptibility (Ladeira & Viegas, 2016; World Health Organization, 1993). Urinary 1-hydroxypyrene, PAH-DNA and PAH-protein adducts are frequently used as a biological marker of PAHs exposure (Jeng & Pan, 2014; Jongeneelen, 2001; Li et al., 2017). Besides, biomarker of effects such as DNA damage,

sister chromatid exchanges, chromosomal aberrations and micronuclei are widely used to assess the association of PAHs exposure and genotoxic effects in human (Ladeira & Viegas, 2016; Zare Jeddi et al., 2021).

Apart from the carcinogenicity of PAHs, the factor of interindividual variation may also lead to adverse health effects. The metabolic pathways of PAHs involved genes encoding xenobiotic metabolizing enzymes such as phase I cytochrome P450s; cytochrome P450 1A1 (*CYP1A1*), and phase II detoxifying enzymes; glutathione S-transferase Mu 1 (*GSTM1*) and glutathione S-transferase Theta 1 (*GSTT1*), which involve in the activation and elimination of carcinogenic PAHs (Sreeja et al., 2005). Enzyme coding genes were considered a risk factor in contributing to changes in the internal dose of PAHs exposure and causing alteration of normal cell function such as DNA damage. Although individuals may be exposed to similar environmental exposures, genetic differences in metabolism may result in markedly different doses at the target site and thus in a different response (World Health Organization, 1993). Biological differences or genetic polymorphisms may exist that cause some individuals to be more susceptible to environmental diseases and serve as markers of susceptibility (USEPA, 2004). Figure 1.1 shows a series of biomarkers which can be applied to enhance the understanding on the entire disease continuum, from exposure to effect, or disease outcome (USEPA, 2004).

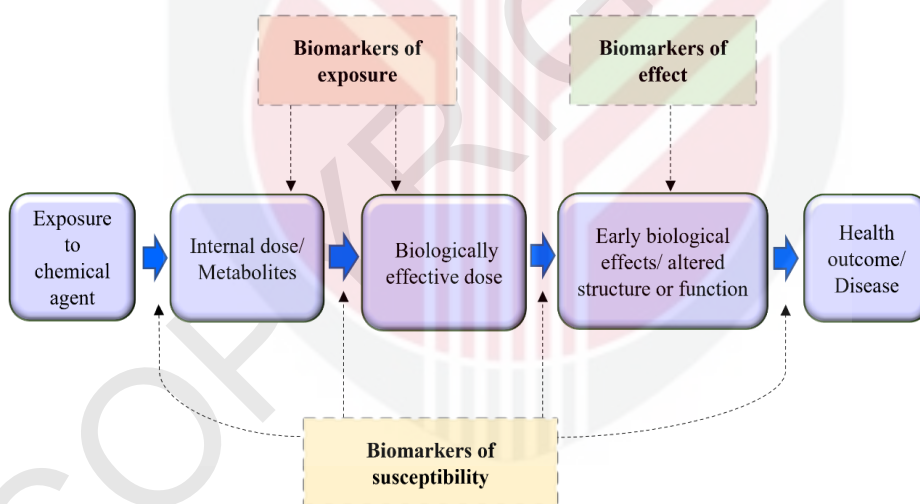


Figure 1.1: Exposure-disease Continuum Involving Biomarkers of Susceptibility, Exposure and Effects [Adapted from USEPA, 2004]

1.2 Problem Statement

The rapid urbanisation, industrial activities, population growth and high traffic volume have resulted in the Klang Valley being constantly exposed to air quality problems. Recent studies in the city of Kuala Lumpur, Malaysia have shown that PM pollution in

this area is greatly contributed by vehicular emissions (Azhari et al., 2021; Suhaimi et al., 2021). According to the Ministry of Transport Malaysia (2019) data, Klang Valley ranked first in the total number of motor vehicles by state in 2019. In addition, the number of newly registered vehicles in Kuala Lumpur increased by 10.95% from 109,365 units in 2018 to 121,340 units in 2019 (Ministry of Transport Malaysia, 2019). Air pollution in the city of Klang Valley is expected to increase as most cities in Malaysia are dominated by private vehicles such as cars and motorcycles (Figure 1.2), which may increase pollutant emissions.

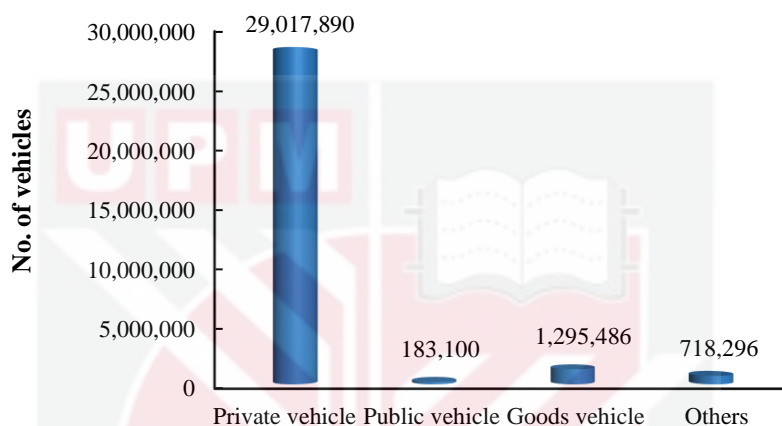


Figure 1.2: Total Motor Vehicles by Type in Malaysia as recorded until 31 Desember 2019 [Adapted from Ministry of Transport Malaysia, 2019].

Malaysia's climate is categorised as an equatorial with uniform temperature and is hot and humid throughout the year (Norazman et al., 2021). Urban schools in Malaysia typically use natural ventilation systems to introduce fresh outdoor air into the classroom as a practical approach to improve the indoor environment. However, the natural ventilation through door and window openings has brought the particles and other air pollutants into the classroom simultaneously. It has been suggested that school located in urban areas with high vehicle densities may contribute to poor indoor air quality in the classrooms (Othman et al., 2022; Zhang et al., 2019). School located nearby heavy traffic area may have high concentrations of outdoor pollutants which can be transferred to the indoor environment via ventilation intake, open doors and windows (Mohammadyan et al., 2017), air exchange rate (Yang et al., 2017) and cracks in the building (Srithawirat et al., 2016). Previous local studies indicated that proximity to traffic affects indoor air quality (IAQ) in classrooms where the outdoor air pollutants could penetrate the indoor environment (Jalaludin et al., 2014; Othman et al., 2019; Suhaimi et al., 2021).

PAHs are associated with carcinogenic effects, among the most important concerns and adverse health effects in humans. The International Agency for Research on Cancer (IARC) classifies benzo[a]pyrene (BaP) as Group 1 (carcinogenic to humans) and

several other PAHs as Group 2A (probably carcinogenic to humans) and 2B (possibly carcinogenic to humans) (IARC, 2021). Exposure to carcinogenic compounds at a young age may pose a significant health risk by causing genetic damage that may increase cancer risk later in life. Previous studies reported that environmental PAHs exposure leads to increased DNA damage in children who attended schools within 500 m of high traffic roads than children who went to rural school in Thailand (Ruchirawat et al., 2007; Tuntawiroon et al., 2007). Genetic damage that occurs early in life may give impact to the risk of carcinogenesis and other chronic diseases in adulthood more than later damage (Zani et al., 2020). Children have a longer lifespan and are more likely to develop chronic diseases that can be triggered by early exposure. Chronic diseases and cancer caused by environmental toxicants are thought to evolve in phases that take years or even decades to develop from their initiation to clinical manifestation. Carcinogenic and toxic exposures in early childhood seem more likely to lead to disease than similar exposures later in life (Mielżyńska et al., 2006; Suk et al., 2003).

Pyrene is one of the major components of PAHs and its urinary metabolites, urinary 1-hydroxypyrene (1-OHP) is the most employed biomarker of exposure to total PAHs (Miri et al., 2018; Sochacka-Tatara et al., 2018; Urbancova et al., 2017). Available evidence showed that children in the youngest age group (6-11 years) had nearly 30% higher urinary 1-OHP levels than adults under the same conditions, suggesting that children appear to be more susceptible to PAHs and have a higher potential health risk (Huang et al., 2006). Changes in the internal PAHs metabolites and the occurrence of DNA damage have the potential to reflect the main effects of PAHs exposure and genetic susceptibility. Apart from the PAHs exposure, genetic polymorphism may be one of the factors leading to differences in internal dose of PAHs excreted, DNA damage and further increased cancer risk. The study of genetic polymorphisms of enzymes that activate and detoxify PAHs helps to identify the individuals who are susceptible to PAHs toxicity.

To the best of our knowledge, none of the study has been conducted on the gene-environment interactions in relation to environmental exposure to PAHs in Malaysia. To date, only three studies have been conducted on the effect of genetic polymorphisms on biomarker of exposure and effects among children population in other countries. However, understanding of the gene-environment interaction for the ambient levels of PAHs exposure among children population is still limited due to inconsistencies findings. In a study by Mielzynska-Svach et al. (2013), no effect of *GSTM1* polymorphism on the level of 1-OHP concentration was found in Polish children, whereas in Mexican children with the *GSTM1* null polymorphism, the risk of high urinary 1-OHP concentrations was five times higher (Sánchez-Guerra et al., 2012). The inconsistencies in the previous literatures seem to indicate that further studies are needed to extend the knowledge of gene-environment interactions in relation to environmental exposure to PAHs, particularly in Malaysian children. In addition, *CYP1A1*, *GSTM1* and *GSTT1* polymorphisms and their modifying effects on children have not been fully elucidated, especially at low levels of PAHs exposure (Nerurkar et al., 2000). Genetic polymorphisms may have different modulation effects at low doses than at high doses (Vineis et al., 2000).

1.3 Study Justification

Investigation on PAHs in the school environment can serve as a baseline data on microenvironmental exposure to PAHs for children in urban traffic area, Klang Valley. Educational buildings like preschools, primary or elementary schools are some of the most important buildings where children typically spend up to 1/3 of their time (around 7–8 h) at these places (Oliveira et al., 2017). According to the current school list from the Ministry of Education Malaysia, 5895 (57.7%) of the 10,218 national primary and secondary schools in Malaysia are located in urban areas. Of these, 851 (90.5%) out of 940 schools are located in urban areas in Selangor and 295 schools (100.0%) are located in Kuala Lumpur. Given these figures for schools in urban areas, special attention should be given to air quality in school buildings, especially for schools located near busy roads. Therefore, understanding the exposure to health-relevant air pollutants in schools has become a priority to the scientific community, in line with the fourth target of the Sustainable Development Goals (SDGs), "excellent education," which requires a focus on educational buildings (Jasso-Pineda et al., 2015; Ochoa-Martinez et al., 2016; Oliveira et al., 2017).

To assess the health risks resulting from exposure to PAHs, it is also necessary to quantify the actual dose entering the body and the resulting changes using various biomarkers (Tuntawiroon et al., 2007). Incorporating biomarker assessment into studies is particularly important in children because the absorbed dose at a given external exposure may be very different from adult due to differences in metabolism, physiology and behaviours (hand-to-mouth activity, playing on the floor) (Oliveira et al., 2017). Previous studies on PAHs exposure in the Klang Valley mainly focused on calculating incremental lifetime cancer risk to estimate the potential health risk of PAHs exposure to the general population (Jamhari et al., 2014; Khan et al., 2015; Omar et al., 2002; Sulong et al., 2019; Suradi et al., 2021). The present study was conducted to fill the knowledge gap on traffic PAHs emission and health effect concerning DNA damage among children living near heavy traffic areas. The concentrations of internal PAHs metabolites (urinary 1-OHP) and DNA damage (assessed by tail moment) were evaluated to investigate the influence of low PAHs concentrations on both biomarkers.

Several epidemiological studies in other countries have reported the association of PAHs exposure with biomarkers of exposure and effects in children population. However, the susceptibility markers involved in PAHs metabolism is still less known, especially among children population, with the fluctuation results were obtained. Therefore, the extension information is needed, especially among the population of Malaysian children. It is important to note that genetic background influence an individual's response to environmental pollutants; thus, studying polymorphisms in human populations is necessary for correct interpretations. The combination assessment of the biomarker of early effects, exposure and susceptibility represent a valuable tool for assessing the potential health effects in the exposed subjects which will provide a comprehensive characterisation of human exposure to PAHs (Alhamdow et al., 2017; Barth et al., 2017; Oliveira et al., 2020). Therefore, this study was conducted to extend the knowledge of gene-environment interactions in relation to environmental PAHs exposure particularly in Malaysian children. In addition, the modifying effects of *CYP1A1*, *GSTT1* and *GSTM1*

polymorphisms on PAHs marker exposure (urinary 1-OHP) and DNA damage were investigated to understand the health risk posed by children exposed to environmental PAHs.

Scientific findings on the characterisation of air pollutants can provide useful information for policymakers to develop underlying action and implement mitigation plans to protect the public, especially children. The findings of this study will not just be beneficial to children but also to those communities living near the heavy traffic areas that might be continuously exposed to traffic emission. More importantly, it is intended that the study's findings would aid in the development of a comprehensive action plan for managing air quality in schools. This research also would help in giving the primary prevention of health specifically in promoting awareness among school children, school staff, parents and other school communities on the risk of air pollution on children's health. The mitigation strategies to reduce air pollution in the schools should begin within the nearby community, school management, as well as local authorities to ensure a clean environment for children.

1.4 Conceptual Framework

A conceptual framework is depicted in Figure 1.3.

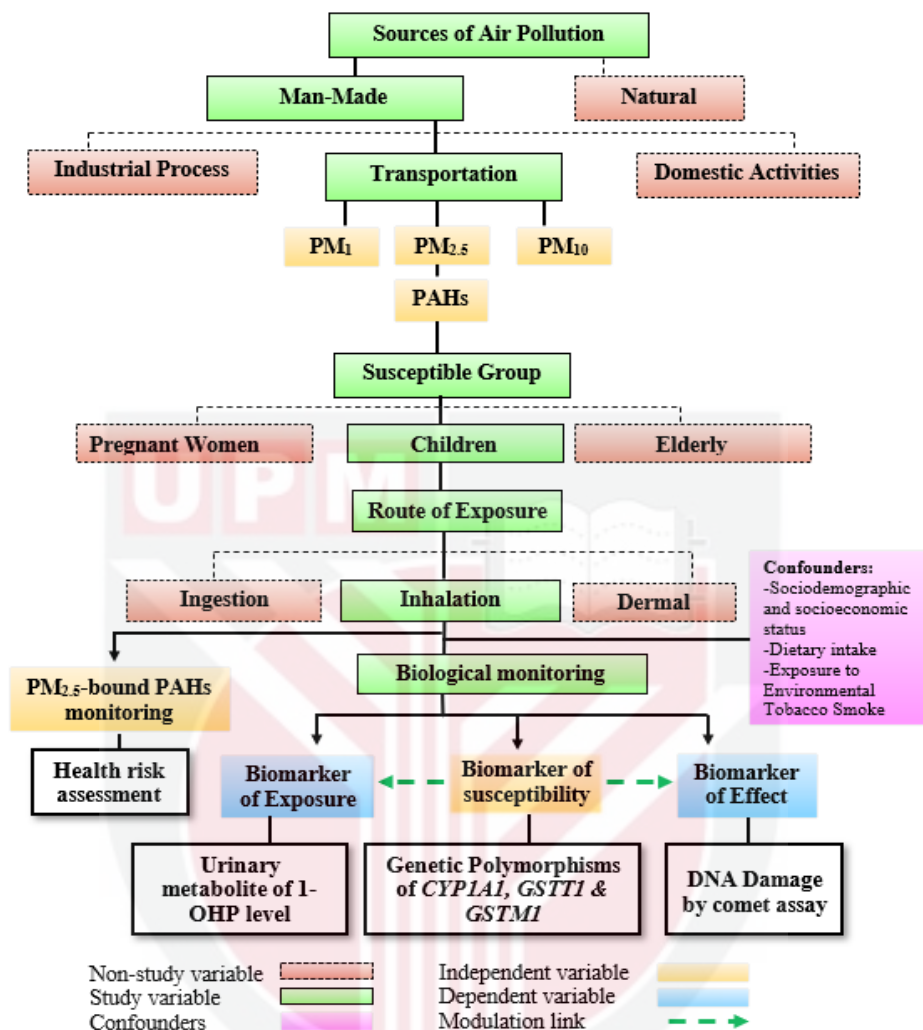


Figure 1.3: Conceptual Framework

This study focused on air pollution emitted from transportation, mainly PM and PAHs in urban city of the Klang Valley. Klang Valley is recognised as one of the most polluted areas in Malaysia. Vehicle emissions significantly contribute to PM pollutants in urban areas. PM_{2.5} is a significant constituent of polluted air and is being intensively studied along with the carcinogenic PAHs bound to it. The particle-associated PAHs are more harmful to human health than gaseous PAHs as they are inhalable and can deposit in the human respiratory system. Route of exposure to PAHs in human occurs via inhalation, ingestion, and dermally from air particles. However, exposure to PAHs from tiny inhalable particles is mainly through inhalation (Suradi et al., 2021). The susceptible population, such as children, is believed to be exposed to a high level of PAHs, especially for those who live and attend schools near high traffic areas. The environmental monitoring of PM and PAHs was carried out among primary school children who were

divided into the high-traffic and low-traffic groups. Environmental monitoring was conducted to evaluate the external doses of PAHs exposed to children. The concentrations of PAHs measured can assist the assessment of carcinogenic risk via health risk assessment using the USEPA approach. In line, biological monitoring was performed to assess the biomarker of exposure, effect and susceptibility with respect to PAHs exposure.

PAHs is rapidly metabolised and conjugated prior to the excretion into urine in the form of urinary metabolites. Urinary 1-OHP is urinary metabolites derived from inhaled PAHs and a specific marker of exposure to PAHs. The genotoxicity impact of exposure to PAHs can be assessed by using biomarker of effects. In this study, the measurement of early biological effects was assessed by evaluating DNA damage using comet assay. The comet assay is reliable method and widely used to evaluate the genotoxic effects of PAHs exposure among children population (Gamboa et al., 2008; Jasso-Pineda et al., 2015; Ruchirawat et al., 2007; Sánchez-Guerra et al., 2012). The biomarker of susceptibility in this study refers to the genetic polymorphisms in individual with respect to exposure to PAHs. This genetic polymorphism could modulate the effect of environmental exposures to genotoxic agents and cause some individuals to be more susceptible to environmentally induced diseases. In this study, the modifying effects of *CYP1A1*, *GSTT1* and *GSTM1* polymorphisms on the biomarker of exposure and effects were investigated. There were linkage studies of *CYP1A1*, *GSTT1* and *GSTM1* polymorphism with the internal dose of 1-OHP and DNA damage (Mielzynska-Svach et al., 2013; Sánchez-Guerra et al., 2012). In this study, confounders that may affect the study's dependent variables such as sociodemographic information, socioeconomic status, dietary intake and exposure to environmental tobacco smoke were also assessed.

1.5 Research Objectives

1.5.1 General Objective

To determine the influence of PAHs and *CYP1A1*, *GSTT1* and *GSTM1* polymorphisms on the PAHs marker exposure (urinary 1-hydroxypyrene) and DNA damage (tail moment) among children in the vicinity of heavy traffic areas in Klang Valley, Malaysia.

1.5.2 Specific Objectives

- i. To determine sociodemographic and socioeconomic information, background of residences, home environment and dietary intake among children in heavy and low traffic areas.
- ii. To evaluate the concentration of PM and PAHs at schools and residences nearby heavy and low traffic areas.
- iii. To evaluate the carcinogenic risk of PAHs by inhalation exposure using the incremental lifetime cancer risk (ILCR) model.

- iv. To compare the concentration of urinary 1-hydroxypyrene (1-OHP), extent of DNA damage and genotype distributions (*CYP1A1*, *GSTT1*, *GSTM1*) among children living nearby heavy and low traffic areas.
- v. To determine the association between PAHs exposure with urinary 1-OHP level and extent of DNA damage among school children.
- vi. To determine the influence of genetic polymorphisms (*CYP1A1*, *GSTT1*, *GSTM1*) on urinary 1-OHP and DNA damage.
- vii. To evaluate the factors that influence the concentration of urinary 1-OHP and DNA damage after controlling the confounders.

1.6 Research Hypothesis

- i. There are significant differences in socio-demographic and socio-economic information, background of residences, home environment and dietary intake between children in heavy and low traffic areas.
- ii. There are significant differences in the concentration of PM and PAHs in schools and residences between heavy and low traffic areas.
- iii. Children in heavy traffic areas had a substantially higher cancer risk than the children in low traffic areas.
- iv. There are significant differences in concentration of urinary 1-OHP, extent of DNA damage and genotype distributions among children between heavy and low traffic areas.
- v. There are significant associations between PAHs exposure with urinary 1-OHP and DNA damage among children.
- vi. There are significant influences of genetic polymorphism (*CYP1A1*, *GSTT1*, *GSTM1*) on urinary 1-OHP and DNA damage among children.
- vii. PAHs exposure is the main factor influencing concentration of urinary 1-OHP and DNA damage among children after controlling the confounders.

1.7 Definition of Variables

1.7.1 Conceptual Definition

1.7.1.1 High and low traffic area

High traffic area are defined as area within 500 m on either side of highways with an Average Daily Traffic (ADT) volume of $\geq 18,000$ vehicles, or within 100 m on either

side of major urban roads with ADT volume of $\geq 15,000$ vehicles (Brauer et al., 2013). Low traffic area refers to area located at 5 km away from highways, major roadways and industrial sites in Selangor.

1.7.1.2 PM₁

PM₁ is a respirable PM with an aerodynamic diameter of 1 μm or less. PM₁ may even pass through the bloodstream (USEPA, 2021).

1.7.1.3 PM_{2.5}

PM_{2.5} is defined as fine inhalable particles with a diameters of 2.5 μm and smaller (USEPA, 2021). PM_{2.5} is more harmful than coarse particles because it could easily penetrate the respiratory tract and reach the circulatory system (Kim et al., 2015).

1.7.1.4 PM₁₀

PM₁₀ is defined as inhalable coarse particulate matter with an aerodynamic diameter of 10 μm or less (USEPA, 2021)

1.7.1.5 Polycyclic aromatic hydrocarbon

Polycyclic aromatic hydrocarbons (PAHs) are persistent organic pollutants (POPs) containing hydrogen and carbon with two or more fused aromatic rings (Abdel-Shafy & Mansour, 2016; Famiyeh et al., 2021). PAHs originate mainly from anthropogenic sources including motor vehicle emissions, fossil fuel burning, coal combustion, oil refining, waste incineration and the coke and steel industries (Khan et al., 2015).

1.7.1.6 Urinary 1-hydroxypyrene

1-hydroxypyrene (1-OHP) is a metabolite of PAH pyrene, a compound that is typically abundant in all PAHs mixtures (Li et al., 2017). Pyrene is one of the main components of PAHs mixture in the air. Pyrene is absorbed into the body by inhalation, skin absorption or ingestion, and then metabolised to 1-OHP, which is ultimately excreted in urine. Urinary 1-OHP has been widely used as a biological marker of PAHs exposure (Jeng & Pan, 2014).

1.7.1.7 DNA damage

DNA damage is an alteration in the chemical structure of DNA that modifies its coding properties and disrupts the cell functions. A major cause of DNA damage is oxidative stress, which is an imbalance between the production of reactive oxygen species (ROS) and an organism's ability to detoxify the reactive intermediates. Environmental factors such as air pollution can induce DNA damage through oxidative stress. DNA strand breaks, which can be detected using a comet assay, are among the main type of DNA damage (Gonzalez-Hunt et al., 2018). DNA damage could serve as a biomarker of early effects of carcinogenic compound exposure.

1.7.1.8 Genetic polymorphism

Genetic polymorphisms are described as variations in DNA sequence that are observed in 1% or more of the population (Norppa, 2004). The biomarkers of susceptibility most frequently used are the genetic polymorphisms of phase I and II metabolizing enzymes (Poblete-Naredo & Albores, 2016). Identifying polymorphisms is crucial in understanding the correlation between genetic variation and susceptibility to toxic exposures (Ramos & Bojang, 2018).

1.7.1.9 *CYP1A1*

The cytochrome P450 (CYP) are the major enzymes involved in phase I metabolism of PAHs. Enzyme CYP P450 1A1, which is encoded by the *CYP1A1* gene is responsible for the metabolic activation of PAHs into reactive metabolites (Stading et al., 2021). There are three genotypes of *CYP1A1* polymorphism wild-type (TT), heterozygous variant (TC) and homozygous variant (CC). The variant of *CYP1A1* genotypes (heterozygous and homozygous) has been associated with elevated induction of the enzyme, and thus increased levels of PAHs reactive metabolites (Xie et al., 2016).

1.7.1.10 *GSTT1* and *GSTM1*

Glutathione s-transferases (GSTs) are key enzymes in phase II metabolism that help the body detoxify various carcinogenic metabolites. GST class theta-1 (*GSTT1*) and GST class mu-1 (*GSTM1*) are the genes encoding for the GST isoenzymes (Wan Rashidi & Bakar, 2019). The deletion polymorphism in *GSTT1* and *GSTM1* (null genotype) is associated with depleted enzyme function, thus increased the toxic metabolites in the body (Abubakar & Sanusi, 2020).

1.7.2 Operational Definition

1.7.2.1 Particulate matter

The real-time sampling of PM₁, PM_{2.5} and PM₁₀ (6 hours) was measured using TSI Dustrak™ DRX Aerosol Monitor 8534, which is based on the light scattering principle. The TSI Dust-TRAK DRX Aerosol Monitor 8534 are battery-operated with data logging that gives a real-time aerosol mass reading. This instrument can measure aerosol concentrations between 0.001 to 150 mg/m³. For 24 h gravimetric measurement of PM_{2.5}, a low volume sampler (Airmetrics MiniVol) was used, with a flow rate of 5.0 L/min

1.7.2.2 Polycyclic aromatic hydrocarbon

PAHs were determined from the filter paper containing PM_{2.5}. The filter paper, Quartz microfibre filters (47 mm diameter size), were digested and cleaned-up using solid-phase extraction (SPE) technique. PAHs samples were quantified using gas chromatography-mass spectrometry (GC-MS). The GC-MS instrument was calibrated with standard mixtures of PAHs.

1.7.2.3 Urinary 1-hydroxypyrene

Urine samples were collected earliest in the morning. Determination of urinary 1-OHP was conducted using solid-phase extraction and high-performance liquid chromatography (HPLC) with fluorescence detection. The concentration of urinary 1-OHP was adjusted with creatinine concentration, and the final concentration was expressed in µmol/mol creatinine.

1.7.2.4 DNA damage

The gold standard for measuring DNA strand breaks is the comet assay (Gonzalez-Hunt et al., 2018). Buccal epithelial cell were used to assess DNA damage using the comet assay. The degree of DNA damage was expressed as tail moment using OpenComet software. Tail moment is defined as the product of tail length and the percentage of DNA in the tail (Hong et al., 2020).

1.7.2.5 Genetic polymorphism

Genomic DNA were extracted from buccal epithelial cells. Genotyping of three PAHs-metabolising enzymes *CYP1A1*, *GSTT1* and *GSTM1* were done by polymerase chain reaction (PCR). *CYP1A1* genotypes were characterised by PCR restriction fragment length polymorphism (PCR-RFLP). *GSTT1* and *GSTM1* null genotypes were determined by conducting multiplex PCR to amplify both genes simultaneously.

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